

NTP RESEARCH REPORT ON IN VIVO REPEAT DOSE BIOLOGICAL POTENCY STUDY OF TRIPHENYL PHOSPHATE (CAS NO. 115-86-6) IN MALE SPRAGUE DAWLEY RATS (HSD: SPRAGUE DAWLEY SD) (GAVAGE STUDIES)

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Peer Review

The draft *NTP Research Report on In Vivo Repeat Dose Biological Potency Study of Triphenyl Phosphate (CAS No. 115-86-6) in Male Sprague Dawley Rats (Hsd: Sprague Dawley SD)* (*Gavage Studies*) was evaluated by the reviewers listed below. These reviewers served as independent scientists, not as representatives of any institution, company, or governmental agency. In this capacity, reviewers determined if the design and conditions of these NTP studies were appropriate and ensured that this NTP Research Report presented the experimental results and conclusions fully and clearly.

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Abstract

Background: Triphenyl phosphate (TPHP) is an organophosphate flame retardant currently on the market and used as a replacement for phased-out polybrominated diphenyl ethers. Toxicity information on this class of chemicals is inadequate. Herein, we use short-term, in vivo transcriptomic studies to inform potential points of departure for TPHP.

Methods: The National Toxicology Program (NTP) conducted a transcriptomic study on TPHP, in male Harlan Sprague Dawley rats. TPHP was dissolved in corn oil. Exposure was once daily for 4 days by oral gavage. TPHP (>99%) was tested at six doses (0, 55, 110, 220, 441, and 881 mg/kg body weight). On Day 5, animals were sacrificed, standard toxicological measures assessed, and the liver taken for gene expression studies using Affymetrix microarrays. Modeling was conducted to identify the benchmark doses (BMDs) associated with the most sensitive apical toxicological endpoints and with transcriptional changes in the liver at a benchmark response of one standard deviation from the mean.

Results: The most sensitive apical endpoints for which BMD values could be obtained were serum HDL (high-density lipoprotein) cholesterol levels, absolute liver weights, relative liver weights, and serum cholesterol levels. The benchmark dose lower confidence limit BMD_Ls (and BMDs) were 39 (79), 48 (136), 71 (103), and 90 (142) mg/kg, respectively. Although serum cholinesterase appeared to be a sensitive endpoint (35–70% decrease) at all doses, beginning with 55 mg/kg (the lowest-observed-effect level), its BMD could not be determined due to poor model fit.

Sensitive transcriptional gene set changes by potency included 14 Gene Ontology Biological Processes with BMD median values below the lower limit of extrapolation (<18.3 mg/kg). The most sensitive gene sets for which a reliable estimate of the BMD could be made were cellular polysaccharide biosynthetic process and oligodendrocyte development with median BMD of 19 mg/kg and BMD_L medians of 11 mg/kg for both gene sets. The most potently affected single genes ranking in the top 10 for increased fold expression change respective to control were *Ces2c* and *Cyp2b1* with BMDs <18.3 mg/kg (the lower limit of extrapolation, $3 \times$ lower than the lowest tested dose) and maximal fold increases of 16.3 and 10.5, respectively. The most potently affected single genes ranking in the top 10 for decreased fold expression change respective to control were control were *Scd* and *G6pc* with BMD of <18.3 mg/kg and BMD_L of 16 mg/kg, respectively, and maximal fold decreases of -11.1 and -5.1, respectively.

Conclusion: Taken together, the most sensitive BMD gene set medians that could be reported occurred at 19 mg/kg (BMD_L 11 mg/kg) and the most sensitive apical endpoint was increased serum HDL cholesterol, with a BMD_L of 39 mg/kg. Cholinesterase inhibition was observed, with effects occurring at all doses including the lowest tested dose of 55 mg/kg. Future studies investigating lower doses would be helpful to obtain more accurate estimates of BMD values for the most sensitively affected genes and for cholinesterase inhibition.

Background

Polybrominated diphenyl ether (PBDE) flame retardant chemicals were voluntarily phased out in 2005 due to concerns about their environmental persistence, bioaccumulation, and association with several adverse human health effects including altered circulating hormone levels, decreased fertility, and impaired neurodevelopment¹⁻³. They have been replaced by organophosphate flame retardants (OPFRs) that are increasingly used in commerce. Like PBDEs, OPFRs can leach from treated materials and persist in the environment. They have been detected in indoor air, household dust, wastewater treatment plant effluent, drinking water, and wildlife samples⁴⁻⁷. They also have been detected in human tissues at levels similar to those of PBDEs⁸. Information is lacking on the incidence and potency of health effects associated with exposure to this chemical class. Furthermore, OPFRs are manufactured as isomeric and commercial mixtures with other non-OPFR flame retardants.

Reported here are the results of a repeat dose study performed in male rats. The goal of this study is to provide a rapid assessment of in vivo biological potency by evaluating a combination of traditional toxicological endpoints and transcriptomics analysis to broadly query biological space for any dose-related change. The justification for using this type of assessment relates to the observation that gene set benchmark dose (BMD) values from short-term transcriptomic studies have been shown to approximate dose responsiveness of the most sensitive apical endpoints from resource intensive guideline toxicological assessments (e.g., carcinogenicity)^{9; 10}. Importantly, the study reported here is not intended to assess or identify hazards. In particular, any observations related to traditional toxicological hazards that are gleaned from qualitative interpretation of the transcriptomics data should be considered hypotheses requiring further evaluation.

This report presents the study results for triphenyl phosphate (TPHP), one of the OPFRs nominated to NTP. Reports on additional OPFRs, once completed, will be published on the NTP website.

Materials and Methods

Study Design

Male Hsd: Sprague Dawley SD rats were obtained from Harlan Laboratories (now Envigo, Inc) (Indianapolis, IN). Males were selected because of the historical precedent of using males in transcriptomic studies to avoid challenges with hormonal cyclicity in female rats that can affect interpretation of gene expression data. On receipt, the rats were 7–8 weeks of age. Animals were quarantined for 7 days, and then randomly assigned to one of five dose groups, each containing five rats. The rats in each dose group then were administered TPHP by gavage in corn oil at a dose level of 0, 55, 110, 220, 441, or 881 mg/kg body weight. These doses correspond to molar equivalencies of 0, 0.169, 0.338, 0.675, 1.35, and 2.7 mmol/kg. Corn oil was selected as the vehicle based on physical chemical properties that indicated the test article would exhibit maximal solubility in corn oil relative to other commonly used vehicles. Dosing of the test article occurred on 4 consecutive days. Dosage volume was 5 mL/kg body weight and was based on the most recently measured body weight. Euthanasia, blood/serum collection, and tissue sample collection were completed on the day following the final administration of the test article (Day 5). Animal identification numbers and a summary of collected endpoints for each animal are presented in Appendix A.

Dose Selection Rationale

Doses were selected based on comparison to NTP subchronic studies of tricresyl phosphate, a chemical structurally similar to TPHP. At dose levels of approximately 1000 mg/kg/day, tricresyl phosphate produced significant histopathological manifestations in the liver of rats after 90 days, which indicated the animals were adequately challenged. An equimolar dose of TPHP was estimated to be approximately 881 mg/kg/day; thus, this dose was selected as the highest dose in the present study.

Chemistry

Procurement, Characterization, and Formulation

Triphenyl phosphate (TPHP; CAS No. 115-86-6; $C_{18}H_{15}O_4P$; molar mass 326.29 g/mol) was obtained from Acros Organics (Geel, Belgium) in three lots (A0293313, A0299574, A0321681). Lots were combined to form a single lot (A8609-1NP), characterized, and formulated by MRI Global (Kansas City, MO).

The identity of the combined lot was confirmed using Fourier transform infrared spectroscopy and 1H and 13C nuclear magnetic resonance spectroscopy. Purity was determined by gas chromatography (GC) with flame ionization detection (FID) as >99%. Dose formulations were prepared in corn oil at target concentrations of 0 (vehicle), 0.0338, 0.0676, 0.135, 0.270, and 0.540 mmol/mL, analyzed by GC-FID, and shipped to Alion (Alion, Research Triangle Park, NC). All formulations were within $\pm 10\%$ of target concentrations. The stability of the corn oil formulations was assessed using the 0.0338-mmol/mL concentration for up to 21 days when stored at ambient temperature in sealed glass bottles under inert gas; the actual concentration was

within 10% of the nominal target concentration on Day 0, demonstrating the stability during the period of use. A summary of the analytical results is presented in Appendix B.

Clinical Examinations and Sample Collection

Clinical Observations

Standard clinical observations were performed within 4 hours post dosing on all study days. Animals were observed for signs of cholinesterase inhibition with specific signs recorded, such as weakness, lethargy, tremors, eye-bulging, salivation, lacrimation, and diarrhea. Animals were weighed on the first day of exposure and on the day of necropsy.

Clinical Pathology

Animals were terminated in random order by CO₂/O₂ (70/30) anesthesia one day after the final day of exposure. Blood samples were taken via cardiocentesis. Five mL of blood was collected into a tube void of anticoagulant and the serum harvested for clinical chemistry, free thyroxine (T4), and cholinesterase measurements. The following clinical chemistry parameters were measured on an Olympus AU400e chemistry analyzer (Olympus America, Inc., Irvin, TX) using reagents obtained from Beckman Coulter (Brea, CA) or Diazyme (Poway, CA): urea nitrogen, creatinine, total protein, albumin, sorbitol dehydrogenase, alanine aminotransferase, aspartate aminotransferase, bile acids, cholesterol, triglycerides, low-density lipoprotein (LDL) cholesterol, HDL cholesterol, and cholinesterase. Free T4 was measured using an MP Biomedical T4 radioimmunoassay kit with an Apex automatic gamma counter (ICN Micromedic Systems, Inc., Huntsville, AL). Toxicological study data tables are presented in Appendix C.

Transcriptomics

Sample Collection for Transcriptomics

Liver transcriptomics were performed on samples taken from three animals per dose group (randomly selected). During necropsy, the entire liver was removed, and liver weight was recorded for each animal. Half the left liver lobe was processed for RNA (ribonucleic acid) isolation. Specifically, three pieces (3-mm cubes) were dissected and transferred to a weigh boat containing liquid nitrogen. Once flash frozen, the liver tissue for each animal was placed into a single, prechilled 2-mL cryotube and stored at or below -70° C. Frozen liver samples were shipped to the Battelle Biomedical Research Center (West Jefferson, OH) on dry ice.

RNA Isolation and cDNA Synthesis

The frozen liver tissues were submerged in 10 volumes of prechilled RNA*later*[®]-ICE (Life Technologies, Carlsbad, CA) and stored at $-20^{\circ}C \pm 10^{\circ}C$ for a minimum of 16 hours. The tissues were removed from the RNA*later*[®]-ICE and weighed. Each liver tissue sample, weighing between 21 and 30 mg, was added to lysis buffer and homogenized using plastic disposable pestles (Fisher Scientific, Pittsburgh, PA). Following homogenization, samples were stored at $-70^{\circ}C \pm 10^{\circ}C$ until RNA was isolated. Samples were thawed and centrifuged. RNA was extracted from the supernatant, subjected to DNase digestion, and isolated using the Qiagen RNeasy Mini Kit (Cat #: 74104; Qiagen, Valencia, CA). Each RNA sample was analyzed for quantity and purity by UV analysis using a NanoDrop ND-1000 Spectrophotometer (NanoDrop

Technologies, Wilmington, DE). All samples were evaluated for RNA integrity using an RNA 6000 Nano Kit with an Agilent 2100 Bioanalyzer (Agilent, Santa Clara, CA) and then stored at $-70^{\circ}C \pm 10^{\circ}C$ until further processing.

Total RNA (100 ng), isolated from each liver sample, was used to synthesize single-stranded DNA, which was subsequently converted into a double-stranded complementary DNA (cDNA) template for transcription. An in vitro transcription (IVT) reaction, which incorporates biotinylated ribonucleotide analogs, then was used to create labeled amplified RNA (aRNA). This RNA target preparation was performed using the Affymetrix GeneChip[®] 3' IVT Express Kit (Cat #: 901228; Affymetrix Inc., Santa Clara, CA) and an Eppendorf Mastercycler[®] thermal cycler (Eppendorf, Hamburg, Germany).

Labeled aRNA was fragmented and subsequently hybridized to the Affymetrix Rat Genome 230 2.0 Array (Cat #: 900505; 31,099 probe sets) using an Affymetrix GeneChip[®] Hybridization Oven 645. The arrays were washed and stained using the Affymetrix GeneChip[®] Hybridization Wash and Stain kit (Cat #: 900720) and a Fluidics Station 450 according to the Affymetrix-recommended protocol (FS450_0001). After washing and staining, arrays were scanned using an Affymetrix GeneChip[®] Scanner 3000 7G, and the raw microarray data (.CEL files) were acquired using Affymetrix GeneChip[®] Command Console[®] Software. The Rat Genome 230 2.0 Array provides coverage of more than 30,000 known transcripts; although the array provides cover the entirety of the rat transcriptome.

Analysis of GeneChip Data Quality

Quality control (QC) measurements were evaluated to determine if the data generated from each Affymetrix GeneChip[®] array were of sufficient quality. Affymetrix-recommended guidelines for evaluating quality were used to evaluate the output files for each GeneChip[®] array using the R/Bioconductor package, Simpleaffy¹¹. The following QC parameters were evaluated for each array: average background, scale factor, percentage of genes scored as present, 3' to 5' ratios for the internal control genes beta-actin and glyceraldehyde-3-phosphate dehydrogenase, values for hybridization control transcripts, and values for poly (A) controls.

For samples that failed to pass QC evaluation due to insufficient data quality, an additional round of RNA isolation and cDNA synthesis was performed and additional GeneChip[®] arrays were run, which were designated with –R after each sample number.

Data Analysis

Statistical Analysis of Body Weights, Organ Weights, Clinical Chemistry

Two approaches were employed to assess the significance of pairwise comparisons between dosed and control groups in the analysis of continuous variables. Organ and body weight data, which have approximately normal distributions, were analyzed using the parametric multiple comparison procedures of Williams^{12; 13} and Dunnett¹⁴. Hormone data and clinical chemistry, which typically have skewed distributions, were analyzed using the nonparametric multiple comparison methods of Shirley¹⁵ and Dunn¹⁶. Jonckheere's test¹⁷ was used to assess the significance of dose-response trends and determine whether a trend-sensitive test (the Williams' or Shirley's test) was more appropriate for pairwise comparisons than a test that assumes no

monotonic dose response (Dunnett's or Dunn's test). Trend-sensitive tests were used when Jonckheere's test was significant at p < 0.01.

Prior to analysis, values identified by the outlier test of Dixon and Massey¹⁸ were examined by NTP staff. Values from animals suspected of illness due to causes other than experimental treatment and values the laboratory indicated as inadequate due to measurement problems were eliminated from the analysis.

Benchmark Dose (BMD) Analysis of Organ Weights and Clinical Pathology

Clinical chemistry, body weight, and organ weight endpoints, which exhibited a significant trend test, were submitted in batch for automated BMD modeling analysis. BMD modeling and analysis was conducted using Benchmark Dose Modeling Software (BMDS) version 2.7. Datasets were executed using the Python BMDS interface (<u>https://pypi.python.org/pypi/bmds;</u> version 0.11), which allows for batch processing of multiple datasets. Data for all endpoints submitted were continuous. A default benchmark response (BMR) of 1 standard deviation (relative to control) was used for all datasets. The following BMDS 2.7 models were used to model the means of the datasets:

- Linear, polynomial
- Power
- Hill
- Exponential M2, M3, M4, M5

Multiple versions of the polynomial model were executed, using the degree of polynomial value equal to 2 to the number of dose groups minus 1 (e.g., if a dataset had five dose groups, a 2° , 3° , and 4° polynomial model would be executed). Models were initialized using BMDS 2.7 model defaults, including restricting the power parameter of the power model and n-parameter of the Hill model to >1 and the beta parameters of the polynomial model to positive or negative, depending on the mean response direction of the dataset. For all models, either a constant or non-constant variance model was selected based on criteria summarized in Table E-1 and outlined in EPA BMD technical guidance¹⁹ and the BMDS 2.7 software.

After model execution, BMDs were selected using the model recommendation procedures generally described¹⁹, and the automated decision logic described in Wignall et al.²⁰ and summarized in Appendix E, Table E-1. Models are placed into one of three possible bins, depending on the results and the bin recommendation logic:

- 1. Failure: model did not successfully complete
- 2. Nonviable model: model successfully completed but with serious issues
- 3. Viable model: candidate for recommended model but with possible warnings

If only one model was in the viable model bin, it was selected as the best-fitting model. If the viable bin had more than one model, consistent with U.S. Environmental Protection Agency (EPA) guidance¹⁹, either the model with the lowest Akaike information criterion (AIC) or lowest benchmark dose lower confidence limit (BMD_L) was selected. If the range of BMD_L values was sufficiently close (<3-fold different), the AIC value was used; otherwise, the BMD_L value was

used. If no model was recommended, no BMD was reported. Details on the analysis criteria and decision tree are provided in Appendix E, Table E-1, and Figure E-2 respectively.

BMD Analysis of Transcriptomics Data

The BMD analysis of the transcriptomic data was performed in accordance with NTP best practices for genomic dose-response modeling as reviewed by an independent panel of experts in October, 2017. These recommendations are described in the 2018 publication National Toxicology Program Approach to Genomic Dose Response Modeling²¹.

Probe set intensities from raw microarray data (CEL files from Affymetrix Rat Genome 230 2.0 Arrays) were normalized by applying the Robust Multi-array Average (RMA) algorithm from the genomics analysis tool, GeneSpring GX 12.6 (Agilent Technology, Foster City, CA). The microarray studies of multiple organophosphate phosphates (data to be reported elsewhere) were performed at the same time such that .CEL files from those related studies were normalized together. Principal component analysis (PCA) of the primary RMA-normalized data indicated a batch effect; due to randomization of the samples in the processing and detailed metadata capture, the source of the batch effect could be identified as the hybridization date. To correct the batch effect, the primary normalized data were loaded into Partek Genomic Suite version 6.16.0812 (St. Louis, MO) and annotated with chemical treatment/dose group and hybridization date annotations. The ANOVA-based remove batch effect function in Partek Genomic Suite then was used to remove quantitative impacts on the hybridization date batch effect. Quality control of the batch-corrected, normalized data was performed by visual inspection, using a PCA plot and normalized intensity histograms (Appendix D).

Dose-response analyses of RMA-normalized, batch-corrected probe set intensities from the TPHP study samples were performed using BMDExpress 2.20.0148 beta²² (https://github.com/auerbachs/BMDExpress-2/releases), an updated version of BMDExpress 1.41 that uses an updated modeling approach. First, control genes (AFFX-) were removed from each data set. A trend test (Williams' trend test^{12; 13}, p < 0.05) and fold change filter (1.5-fold change up or down relative to vehicle control for probe sets) was applied to each data set to remove probe sets demonstrating no response to chemical treatment from subsequent analysis. These filter criteria were empirically determined, with the goal of balancing false discovery with reproducibility. The criteria are consistent with the MicroArray Quality Control recommendations to combine the nominal p-value threshold with a fold-change filter to maximize replicability of transcriptomic findings across labs. Hill, power, linear, polynomial 2°, exponential 2, exponential 3, exponential 4, and exponential 5 dose-response models then were fit to the probe sets that passed the trend test and fold change filter. All gene expression data analyzed in BMDExpress were log2 transformed, and thus nearly all probe sets exhibit constant variance across the doses. For this reason and for efficiency purposes, each model was run assuming constant variance. Lacking any broadly applicable guidance regarding the level of change in gene expression that is considered to be biologically significant, a BMR of $1 \times$ standard deviation was used in this study. This approach enables standardization of the BMR between apical endpoints and transcriptomic endpoints and provides a standard for use across multiple chemicals tested in this rapid screening paradigm. The expression direction (up- or down-regulated) for each probe set was determined by a trend test intrinsic to the model executables contained in BMDExpress provided by EPA.

To identify the best-fit model for each fitted probe set, the AIC for each fitted model was compared and the model with the lowest AIC selected. The best model for each probe set was used to calculate the BMD, BMD_L, and BMD upper confidence limit (BMD_U). The specific parameter settings, selected from the BMDExpress software when performing probe set-level BMD analysis, were as follows: maximum iterations - 250, confidence level - 0.95, BMR factor – 1 (the multiplier of the SD that defined the BMD), restrict power – no restriction, and constant variance – selected. The specific model selection setting in the BMDExpress software when performing probe set-level BMD analysis was as follows: best poly model test - lowest AIC, flag Hill model with 'k' parameters - <1/3 the lowest positive dose, and best model selection with flagged Hill model - include flagged Hill model. The inclusion of the flagged models is deviation from EPA guidance. The justification for this deviation relates to subsequent use of the data in which the probe set BMD values are grouped into gene sets from which a median BMD is derived. If the probe sets were removed from the analysis or forced to another model, the probe set might not be counted in the gene set analysis and could lead to loss of "active" gene sets. Importantly, most of the probe sets that produce flagged Hill models show highly potent responses and should therefore be counted in the analysis. Probe-level data ranked by fold change are reported; probe sets were removed according to the following criteria: global goodness-of-fit p-value <0.1 and BMD_U/BMD_L ratio >40.

To perform Gene Ontology (GO Annotation accession date: 10/7/2017) gene set analysis, only GO terms with ≥ 10 or ≤ 250 annotated genes measured on the gene expression platform were considered. Before populating the GO terms, the best-fit model for each probe set was subject to a filtering process to remove those probe sets (1) with a BMD >highest tested dose, (2) that mapped to more than one gene, (3) that had a global goodness-of-fit p-value <0.1, and (4) with a BMD_U/BMD_L ratio >40. GO terms that were at least 5% populated and contained 3 genes that passed the above criteria were considered "active" (i.e., responsive to chemical treatment). For reporting in the body of the manuscript, GO terms populated with identical sets of differentially expressed genes were filtered to limit redundancy in reporting based on the following selection criteria: (1) highest percentage populated and (2) most specific/highest GO level. Redundant GO terms failing to differentiate based on these criteria were retained and reported. A complete list of "active" GO terms is included in supplemental material accessible online at https://doi.org/10.22427/NTP-DATA-RR-8. Finally, to avoid effects of model extrapolation, GO terms exhibiting a median BMD 3-fold less than the lowest positive dose (due, in some cases, to inclusion of flagged Hill models, as noted above) were assigned a default value threshold value and no BMD_L or BMD_U value was reported. A summary of the BMDExpress gene expression analysis pipeline used in this study is shown in Appendix E, Figure E-2.

Data Accessibility

Primary and analyzed data used this study are available to the public at <u>https://doi.org/10.22427/NTP-DATA-RR-8</u>.

Results

Animal Condition, Body and Organ Weights

All animals survived to the end of the study. Final mean body weight was significantly decreased in the highest dose group as compared to the vehicle control group (Table 1). The BMD (BMD_L) for terminal body weight at study Day 4 (SD4) was 486 (278) mg/kg. Assessment of clinical parameters in a subset of animals found sporadic effects, including red nasal discharge and loose stools with no difference in incidence between vehicle controls and those animals exposed to TPHP (data not shown).

Table 1. Mean Body Weight Summary

Study Day (SD)	0 mg/kg N = 5	55 mg/kg N = 5	110 mg/kg N = 5	220 mg/kg N = 5	441 mg/kg N = 5	881 mg/kg N = 5	BMD _{1Std} (mg/kg)	BMD _{L1Std} (mg/kg)
SD0 (g)	258.9 ± 3.6	256.6 ± 5.4	253.5 ± 3.8	261.5 ± 4.2	254.0 ± 6.7	247.8 ± 4.5	NST	NST
SD4 (g)	$276.3 \pm 4.6^{**}$	271.3 ± 5.1	271.9 ± 3.7	278.8 ± 4.7	262.4 ± 8.1	$232.5\pm9.4^{**}$	486	278

Data are displayed as mean \pm standard error of the mean.

Statistical analysis performed by Jonckheere's (trend) and Williams' or Dunnett's (pairwise) tests.

Statistical significance for the control group indicates a significant trend test.

**Statistically significant at $p \le 0.01$.

Benchmark response (BMR) set at 1 standard deviation from the mean.

SD0 = Study Day 0, the first day of dosing; SD4 = Study Day 4, the day of necropsy.

NST = BMD modeling not conducted due to nonsignificant trend test.

At necropsy, a dose-dependent increase in absolute and relative liver weights and relative brain weight occurred (Table 2). The BMD (BMD_L) for increased absolute liver weight was 136 (48) mg/kg and for relative liver weight was 103 (71) mg/kg. Of note is that the change in relative brain weight is an artifact of significant dose-dependent decreases in body weight. The absolute brain weight was not affected by chemical exposure (Table 2).

Clinical Chemistry

Among the clinical chemistry endpoints evaluated, the most sensitive to TPHP exposure were increased serum levels of HDL cholesterol and total cholesterol with BMD (BMD_L) of 79 (39) and 142 (90) mg/kg, respectively (Table 3). LDL cholesterol levels were also increased in this study, although at a relatively higher dose [BMD (BMD_L) = 213 (123) mg/kg].

Albumin concentrations were decreased in the high-dose group (p < 0.05) and exhibited a BMD (BMD_L) of 576 (322) mg/kg. Globulin concentrations had a statistically significant trend increase, but not a significant pairwise change from the vehicle control group at any specific dose level (trend; p < 0.05). Globulin levels exhibited a BMD (BMD_L) of 328 (174) mg/kg. The combination of these changes resulted in a significant decrease in the albumin:globulin ratio for the 220 mg/kg and higher dose groups. The BMD (BMD_L) for the albumin:globulin ratio was 147 (103) mg/kg.

Endpoint	0 mg/kg N = 5	55 mg/kg N = 5	110 mg/kg N = 5	220 mg/kg N = 5	441 mg/kg N = 4–5	881 mg/kg N = 5	BMD _{1Std} (mg/kg)	BMD _{L1Std} (mg/kg)
Terminal Body Weight (SD4) (g)	$276.3 \pm 4.6 **$	271.3 ± 5.1	271.9 ± 3.7	278.8 ± 4.7	262.4 ± 8.1	232.5 ± 9.4**	486	278
Brain Weight Absolute (g)	1.73 ± 0.05	1.74 ± 0.03	1.77 ± 0.04	1.81 ± 0.03	1.81 ± 0.03	1.78 ± 0.01	NST	NST
Brain Weight Relative (mg/g)	6.28 ± 0.23**	6.43 ± 0.06	6.52 ± 0.13	6.48 ± 0.03	$7.03\pm0.14*$	$7.73 \pm 0.37 **$	PMF	PMF
Liver Weight Absolute (g)	$11.46 \pm 0.19 **$	11.83 ± 0.56	11.84 ± 0.29	$12.95 \pm 0.48*$	$12.92 \pm 0.64*$	$13.09 \pm 0.38*$	136	48
Liver Weight Relative (mg/g)	$41.50 \pm 0.58 **$	43.52 ± 1.41	43.51 ± 0.63	46.42 ± 1.28	$49.14 \pm 1.00 **$	56.74 ± 3.22**	103	71

Table 2. Organ Weights Summary

Data are displayed as mean \pm standard error of the mean.

Relative organ weights (organ-weight-to-body-weight ratios) are given as mg organ weight/g body weight.

Statistical analysis performed by Jonckheere's (trend) and Williams' or Dunnett's (pairwise) tests.

Statistical significance for the control group indicates a significant trend test.

*Statistically significant at $p \le 0.05$; **statistically significant at $p \le 0.01$.

Benchmark response (BMR) set at 1 standard deviation from the mean.

Statistical significance for a treatment group indicates a significant pairwise test compared to the vehicle control group.

PMF = no BMD/BMDL selected due to poor model fit; NST = BMD modeling not conducted due to nonsignificant trend test.

Endpoint	0 mg/kg N = 5	55 mg/kg N = 5	110 mg/kg N = 5	220 mg/kg N = 5	441 mg/kg N = 5	881 mg/kg N = 5	BMD _{1Std} (mg/kg)	BMD _{L1Std} (mg/kg)
Globulin (g/dL)	$2.5 \pm 0.07 **$	2.56 ± 0.07	2.52 ± 0.05	2.66 ± 0.05	2.76 ± 0.06	2.72 ± 0.15	328	174
A/G Ratio	$1.38 \pm 0.03^{**}$	1.31 ± 0.01	1.34 ± 0.02	$1.28\pm0.02*$	$1.22 \pm 0.03 **$	$1.16 \pm 0.07 ^{**}$	147	103
Albumin (g/dL)	$3.44\pm0.07*$	3.36 ± 0.07	3.38 ± 0.04	3.4 ± 0.04	3.36 ± 0.06	$3.12\pm0.06*$	576	322
Cholesterol (mg/dL)	$101.0 \pm 5.8^{**}$	115.0 ± 6.5	122.4 ± 6.2	$127.8\pm6.3^*$	$145.4 \pm 7.3^{**}$	$170.0 \pm 19.4^{**}$	142	90
LDL Cholesterol (mg/dL)	$21.8\pm0.7^{**}$	21.8 ± 1.1	23.2 ± 0.9	23.8 ± 1.5	23.4 ± 1.1	$32.8\pm6.4*$	213	123
HDL Cholesterol (mg/dL)	$46.0 \pm 2.9^{**}$	54.6 ± 2.7	$56.8\pm2.8*$	$60.6 \pm 2.9 **$	$70.6 \pm 3.2 $ **	$78.2 \pm 5.1 **$	79	39

Table 3. Clinical Chemistry Summary

Data are displayed as mean \pm standard error of the mean.

Statistical analysis performed by Jonckheere's (trend) and Shirley's or Dunn's (pairwise) tests.

Statistical significance for the control group indicates a significant trend test.

*Statistically significant at $p \le 0.05$; **statistically significant at $p \le 0.01$.

Benchmark response (BMR) set at 1 standard deviation from the mean.

Statistical significance for a treatment group indicates a significant pairwise test compared to the vehicle control group.

A/G Ratio = ratio of albumin to globulin.

Hormones and Enzymes

A dose-dependent decrease in free thyroxine was observed, beginning with the 220 mg/kg dose group (p < 0.05; Table 4). The BMD (BMD_L) for free thyroxine was 178 (139) mg/kg. A statistically significant decrease in serum cholinesterase (-36 to 70%) was noted in all dose groups, beginning with the 55 mg/kg group (p < 0.01); however, a BMD (BMD_L) could not be calculated due to the inability to fit the data to an appropriate model. Testing lower doses in future studies will therefore be necessary to calculate a benchmark dose associated with increased cholinesterase in the context of TPHP exposure.

Endpoint	0 mg/kg N = 5	55 mg/kg N = 5	110 mg/kg N = 5	220 mg/kg N = 5	441 mg/kg N = 5	881 mg/kg N = 5	BMD _{1Std} (mg/kg)	BMD _{L1Std} (mg/kg)
Free Thyroxine (μg/dL)	5.122 ± 0.227**	5.114 ± 0.196	$\begin{array}{c} 4.678 \pm \\ 0.304 \end{array}$	$\begin{array}{c} 4.274 \pm \\ 0.263 \ast \end{array}$	$4.066 \pm 0.331*$	$1.870 \pm 0.429 **$	178	139
Cholinesterase (IU/L)	277.4 ± 16.6**	178.4 ± 8.0**	162.6 ± 7.6**	116.6 ± 5.8**	108.8 ± 8.3**	$82.8 \pm 2.1^{**}$	PMF	PMF

Data are displayed as mean \pm standard error of the mean.

Statistical analysis performed by Jonckheere's (trend) and Shirley's or Dunn's (pairwise) tests.

Statistical significance for the control group indicates a significant trend test.

*Statistically significant at $p \le 0.05$; ** statistically significant at $p \le 0.01$.

Benchmark response (BMR) set at 1 standard deviation from the mean.

Statistical significance for a treatment group indicates a significant pairwise test compared to the vehicle control group. $PMF = no BMD/BMD_L$ selected due to poor model fit.

Apical Endpoint and Clinical Chemistry Benchmark Dose Summary Table

A summary of the calculated BMDs for each toxicological endpoint is provided in Table 5. The lowest-observed-effect level (LOEL) and no observed effect level (NOEL) is provided and may be informative for endpoints that lack a calculated BMD due to poor model fit or deficiencies in model performance characteristics.

Table Number	Endpoint	NOEL (mg/kg)	LOEL (mg/kg)	BMD _{L1Std} (mg/kg)	BMD _{1Std} (mg/kg)
3	HDL Cholesterol	55	110	39	79
2	Liver Weight Absolute	110	220	48	136
4	Cholinesterase	ND	55	_	_
2	Liver Weight Relative	220	441	71	103
3	Cholesterol	110	220	90	142
3	A/G Ratio	110	220	103	147
3	LDL Cholesterol	441	881	123	213
4	Free Thyroxine	110	220	139	178
3	Globulin	ND	ND	174	328
2	Terminal Body Weight (SD4)	441	881	278	486
3	Albumin	441	881	322	576
2	Brain Weight Relative	220	441ª	-	-

Table 5. BMD, NOEL, and LOEL Summary for Apical Endpoints, Sorted by LOEL or $BMD_{\rm L}$ from Low to High

^aChange in relative brain weight is an artifact of decreases in overall body weight. The chemical had no effect on absolute brain weight. Benchmark response (BMR) set at 1 standard deviation from the mean. ND = not determined. – indicates value not reported due to poor model performance. Values in bold text indicate the LOEL of endpoints for which a BMD could not be calculated.

Gene Set BMD Analysis

Chemical-induced alterations in liver gene transcript expression were examined to determine those gene sets most sensitively affected by exposure to TPHP. To that end, BMD analysis of transcripts and gene sets (GO Biological Processes) was conducted to determine the potency of the chemical to elicit gene expression changes in the liver. This analysis used transcript-level BMD data to assess an aggregate score of gene set potency (median transcript BMD) and enrichment.

The "active" gene sets with the lowest BMD median value are shown in Table 6. The gene sets in Table 6 should be interpreted with caution from the standpoint of the underlying biology and instead should be considered a metric of potency for chemical-induced transcriptional changes.

Fourteen gene sets had estimated BMD median values below the lower limit of extrapolation (<18.3 mg/kg, or three times smaller than the lowest tested dose). These sensitive gene sets included GO:0090181 Regulation of cholesterol metabolic process and GO:0009914 Hormone transport. The most sensitively affected GO Biological Processes for which a BMD value could be reliably calculated were GO:0033692 Cellular polysaccharide biosynthetic process and GO:0014003 Oligodendrocyte development, each with BMDs of 19 mg/kg and BMD_Ls of 11 mg/kg. The full list of affected gene sets is available at <u>https://doi.org/10.22427/NTP-DATA-RR-8</u>.

Category Name	Input Genes /Platform Genes in Gene Set	% Gene Set Coverage	BMD _{1Std} Median of Gene Set Transcripts (mg/kg)	Median BMD _{L1Std} - BMD _{U1Std} (mg/kg)	Genes with Changed Direction Up	Genes with Changed Direction Down
GO:0002548 monocyte	3/26	12%	<18.3	NR	0	3
chemotaxis						
GO:1904031 positive regulation of cyclin- dependent protein kinase activity	3/35	9%	<18.3	NR	1	2
GO:0045739 positive regulation of DNA repair	5/45	11%	<18.3	NR	3	2
GO:0007095 mitotic G2 DNA damage checkpoint	3/17	18%	<18.3	NR	2	1
GO:0031572 G2 DNA damage checkpoint	3/17	18%	<18.3	NR	2	1
GO:0032330 regulation of chondrocyte differentiation	3/36	8%	<18.3	NR	2	1
GO:0090181 regulation of cholesterol metabolic process	3/27	11%	<18.3	NR	3	0
GO:0060192 negative regulation of lipase activity	3/13	23%	<18.3	NR	1	2
GO:1901264 carbohydrate derivative transport	3/48	6%	<18.3	NR	3	0
GO:0009914 hormone transport	5/91	5%	<18.3	NR	2	3
GO:0001942 hair follicle development	3/35	9%	<18.3	NR	1	2
GO:0044773 mitotic DNA damage checkpoint	4/35	11%	<18.3	NR	3	1
GO:0010972 negative regulation of G2/M transition of mitotic cell cycle	4/35	11%	<18.3	NR	3	1
GO:0042698 ovulation cycle	5/43	12%	<18.3	NR	2	3
GO:0033692 cellular polysaccharide biosynthetic process	3/25	12%	19	11-42	3	0

 Table 6. Top 20 GO Biological Process Gene Sets Ranked by Potency of Perturbation (Sorted by BMD Median)

Category Name	Input Genes /Platform Genes in Gene Set	% Gene Set Coverage	BMD _{1Std} Median of Gene Set Transcripts (mg/kg)	Median BMD _{L1Std} - BMD _{U1Std} (mg/kg)	Genes with Changed Direction Up	Genes with Changed Direction Down
GO:0014003 oligodendrocyte development	3/31	10%	19	11-39	3	0
GO:0060487 lung epithelial cell differentiation	3/28	11%	22	9-64	3	0
GO:0051453 regulation of intracellular pH	4/60	7%	23	6-95	1	3
GO:0030004 cellular monovalent inorganic cation homosetasii	6/79	8%	24	6-98	2	4
GO:0072348 sulfur compound	4/27	15%	25	14-59	1	3

 $\overline{NR} = The BMD_L-BMD_U$ range is not reportable because the BMD median is below the lower limit of extrapolation (less than 1/3 of the lowest tested dose in this study).

Differentially Expressed Gene Analysis

The top 10 up- and down-regulated genes based on maximum fold change relative to control that passed the criteria for BMD estimation (fold change >|1.5| and significant Williams' trend test, global goodness of fit >0.1, and BMD_U/BMD_L <40), are shown in Table 7.

The most potently affected genes, experiencing the greatest maximal fold change increase, were *Ces2c* (carboxylesterase 2c) and *Cyp2b1* (cytochrome P450 2b1) and an unmapped transcript, each with a median BMD below the lower limit of extrapolation. Hepatic transcript expression increased dose-dependently and maximal fold changes for *Ces2c* and *Cyp2b1* occurred at the highest tested dose at 16.3- and 10.5-fold above vehicle exposed animals. The most potently induced genes with maximal fold increase for which a BMD and BMD_L could be calculated reliably were *Abcc3* [also known as multidrug resistance 3 (MDR3), BMD_L = 26 mg/kg, maximal fold increase of 12.7] and *Per2* (period circadian regulator 2, BMD_L = 36 mg/kg, maximal fold increase of 7.2).

The most potently affected genes experiencing the greatest maximal fold change decrease were *Scd* (stearoyl-CoA desaturase) with a median BMD below the lower limit of extrapolation (<18.3 mg/kg; maximal fold change of -11.1) and *G6pc* (glucose-6-phosphatase, catalytic subunit) with a median BMD_L value of 16 mg/kg (maximal fold change of -5.1).

A 66	Carra Carrahal I	0	BMD _{1Std}	Fold	Fold	Fold	Fold	Fold
Allymetrix Droho ID	Gene Symbol and	Gene	(BMD _{L1Std} -BMD _{U1Std})	Change	Change	Change	Change	Change
Probe ID	Anases	Iname	in mg/kg	55 mg/kg	110 mg/kg	220 mg/kg	441 mg/kg	881 mg/kg
↑ Increased t	transcripts with max	ximal fold change vs. vehicle control						
1368718_at	Aldh1a7	aldehyde dehydrogenase family 1, subfamily A7	414 (263–962)	-1.8	4.2	8.6	9.5	69.4
1368569_at	Akr1b7	aldo-keto reductase family 1, member B7	238 (186-329)	-1.4	1.1	2.5	2.9	29.0
1368905_at	Ces2c; LOC100910040	carboxylesterase 2C	<18.3 (NR)	2.3	2.8	4.4	10.3	16.3
1369698_at	Abcc3	ATP binding cassette subfamily C member 3	48 (26–80)	1.3	2.8	5.6	9.8	12.7
1371076_at	Cyp2b1; Cyp2b2; LOC100909962	cytochrome P450, family 2, subfamily b, polypeptide 1	<18.3 (NR)	3.1	5.5	6.9	9.0	10.5
1368303_at	Per2	period circadian regulator 2	48 (36–73)	2.1	2.8	3.3	7.2	4.6
1397924_at	_	_	<18.3 (NR)	2.2	2.9	3.4	6.7	5.8
1371089_at	Gsta3	glutathione S-transferase alpha 3	84 (57–152)	1.1	1.4	2.5	4.6	6.6
1370269_at	Cyplal	cytochrome P450, family 1, subfamily a, polypeptide 1	153 (115–278)	1.2	1.2	1.6	3.0	6.1
1381811_at	-	_	58 (42–94)	1.2	2.2	2.6	5.9	2.1
↓ Decreased	transcripts with ma	ximal fold change vs. vehicle control						
1395403_at	Stac3	SH3 and cysteine rich domain 3	81 (56–143)	-1.4	-1.5	-2.4	-7.0	-11.6
1370355_at	Scd; Scd1	stearoyl-CoA desaturase	<18.3 (NR)	-3.8	-6.9	-6.0	-10.5	-11.1
1386977_at	Car3	carbonic anhydrase 3	201 (141-332)	-1.0	-1.1	-1.6	-2.5	-9.8
1397205_at	Dhrs7; LOC100364391	dehydrogenase/reductase 7	163 (118–255)	1.9	1.2	-1.3	-2.4	-7.7
1390672_at	Rprm	reprimo, TP53 dependent G2 arrest mediator homolog	59 (42–95)	-2.2	-2.2	-2.9	-5.9	-3.3
1371102_x_at	LOC100134871; LOC689064	beta-globin	485 (273-825)	-1.2	-1.0	-1.1	-1.4	-5.4
1368171_at	Lox	lysyl oxidase	208 (122-262)	-1.1	1.3	-2.7	-5.4	-5.2
1370725_a_at	Aoc3; G6pc; Psme3	glucose-6-phosphatase, catalytic subunit	27 (16–53)	-2.0	-2.3	-3.8	-5.1	-4.5
1371143_at	Serpina7	serpin family A member 7	330 (167–546)	-1.1	-1.3	-1.4	-1.5	-5.1
1368172 a at	Lox	lysyl oxidase	210 (119-306)	-1.4	1.1	-2.7	-4.7	-5.1

Table 7. Top Differentially Expressed Probe Sets Ranked by Fold Change

- = A gene name cannot be identified because the probes no longer align to the rat genome (i.e., no perfect sequence matches the most recent rat genome assembly, Rnor_6.0). <18.3 = A best-fit model as identified calculated a BMD that was less than 1/3 of the lowest tested dose in this study.

NR = The BMD_L-BMD_U range is not reportable because the BMD median is below the lower limit of extrapolation (less than 1/3 of the lowest tested dose in this study).

Discussion

As PBDE flame retardants have been phased out due to concerns regarding their toxicity and persistence, the use of replacement OPFRs such as TPHP has increased. Limited toxicity data are available to estimate the potential adverse health effects of OPFRs. This study used a transcriptomic approach and standard toxicological endpoints to determine short-term biological activity of TPHP.

Serum cholinesterase was significantly and markedly decreased for all dose groups and appeared to be a sensitive apical measure. These findings are consistent with several reports that show the classic cholinesterase inhibition in organophosphates including TPHP²³. As the LOEL for the study, cholinesterase inhibition appeared to be the most sensitive apical measure; cholinesterase inhibition was so marked at all doses that a BMD value could not be determined due to poor model fit. Further studies are warranted to assess cholinesterase effects at concentrations <55 mg/kg to obtain an accurate point of departure. The most sensitive apical endpoint for which a BMD could be determined was HDL cholesterol with a BMD_L (BMD) of 39 (79) mg/kg. Dosedependent increases in absolute and relative liver weight [48 (136) mg/kg and 71 (103) mg/kg] and cholesterol [90 (142) mg/kg] for BMD_L (BMD) were the next most sensitive apical endpoint changes.

Not surprisingly, transcriptional changes in the liver following TPHP exposure occurred at dose levels below that for which changes in circulating cholinesterase and cholesterol levels were observed. The most sensitively affected gene sets for which a reliable BMD_L could be estimated were cellular polysaccharide biosynthetic process and oligodendrocyte development, both with a BMD_L median value of 11 mg/kg. Fourteen GO Biological Processes were potently affected below the lower limit of extrapolation from the dose curve (BMD < 18.3 mg/kg). This finding suggests that further testing at doses lower than 55 mg/kg would be useful toward refining estimates of the transcriptional point of departure.

The most potently affected single genes with maximal fold increases in hepatic expression included *Ces2c* (carboxylesterase 2C) and *Cyp2b1* (cytochrome P450 family 2, subfamily b, polypeptide 1), each with median BMD values below the lower limit of extrapolation (<18.3 mg/kg). Prior studies investigating hepatic lipid metabolism and effects of TPHP exposure in mice indicated the potential for TPHP to inhibit activity of carboxylesterases using proteomic approaches; this effect was confirmed using activity assays derived from in vitro mouse crude liver lysates, using 50 μ M TPHP²⁴. Other genes with maximal fold increases and potency estimated below the lowest tested dose were *Abcc3* [also known as MRP3 or multidrug resistance 3 transporter; BMD_L (BMD) of 26 (48) mg/kg] and *Per2* [period circadian regulator 2; BMD_L (BMD) of 36 (48) mg/kg], which each have functions in xenobiotic metabolism regulation and transport.

The most potently affected single genes with maximal fold decreases in hepatic expression included *Scd* (stearoyl-CoA desaturase, an enzyme involved in fatty acid synthesis; BMD < 18.3 mg/kg) and *G6pc* (glucose-6-phosphate catalytic subunit, a key enzyme in maintaining glucose homeostasis; BMD_L of 16 mg/kg).

Because the target organ(s) were unknown, identifying meaningful gene sets from a single organ (e.g., liver) that were correlated with noncancer apical responses was not necessarily expected, as

has been suggested in the literature. This analysis therefore focused on dose-response relationships⁹. Our data indicate the TPHP doses associated with the initiation of molecular transcriptional changes and with alterations in apical endpoints.

Under the conditions of this short-duration transcriptomic study in Harlan Sprague Dawley rats, the most sensitive point of departure with a reliable estimate was $11 \text{ mg/kg} (BMD_L)$ for hepatic transcriptional gene set alterations. Increased serum HDL cholesterol was the most sensitive apical endpoint, with a BMD_L of 39 mg/kg, followed by cholinesterase inhibition, which was marked at all tested doses, starting at 55 mg/kg. Follow-up studies that investigate transcriptional and apical endpoint changes at lower doses will be a useful future direction to better resolve transcriptional changes that occur with the greatest potency.

References

1. Frederiksen M, Thomsen M, Vorkamp K, Knudsen LE. Patterns and concentration levels of polybrominated diphenyl ethers (PBDEs) in placental tissue of women in Denmark. Chemosphere. 2009; 76(11):1464-1469. <u>http://dx.doi.org/10.1016/j.chemosphere.2009.07.017</u>

2. Herbstman JB, Sjodin A, Kurzon M, Lederman SA, Jones RS, Rauh V, Needham LL, Tang D, Niedzwiecki M, Wang RY et al. Prenatal exposure to PBDEs and neurodevelopment. Environ Health Perspect. 2010; 118(5):712-719. <u>http://dx.doi.org/10.1289/ehp.0901340</u>

3. Meeker JD, Stapleton HM. House dust concentrations of organophosphate flame retardants in relation to hormone levels and semen quality parameters. Environ Health Perspect. 2010; 118(3):318-323. <u>http://dx.doi.org/10.1289/ehp.0901332</u>

4. Meeker JD, Cooper EM, Stapleton HM, Hauser R. Urinary metabolites of organophosphate flame retardants: temporal variability and correlations with house dust concentrations. Environmental Health Perspectives. 2013; 121(5):580-585. http://dx.doi.org/10.1289/ehp.1205907

5. Porte C, Barcelo D, Albaiges J. Quantitation of total versus selected polychlorinated biphenyl congeners in marine biota samples. J Chromatogr. 1988; 442:386-393. http://dx.doi.org/10.1016/S0021-9673(00)94488-1

6. Sundkvist AM, Olofsson U, Haglund P. Organophosphorus flame retardants and plasticizers in marine and fresh water biota and in human milk. J Environ Monit. 2010; 12(4):943-951. http://dx.doi.org/10.1039/b921910b

7. van der Veen I, de Boer J. Phosphorus flame retardants: properties, production, environmental occurrence, toxicity and analysis. Chemosphere. 2012; 88(10):1119-1153. http://dx.doi.org/10.1016/j.chemosphere.2012.03.067

8. Hoffman K, Fang M, Horman B, Patisaul HB, Garantziotis S, Birnbaum LS, Stapleton HM. Urinary tetrabromobenzoic acid (TBBA) as a biomarker of exposure to the flame retardant mixture Firemaster(R) 550. Environ Health Perspect. 2014; 122(9):963-969. https://doi.org/10.1289/ehp.1308028

9. Thomas RS, Wesselkamper SC, Wang NC, Zhao QJ, Petersen DD, Lambert JC, Cote I, Yang L, Healy E, Black MB et al. Temporal concordance between apical and transcriptional points of departure for chemical risk assessment. Toxicol Sci. 2013; 134(1):180-194. <u>http://dx.doi.org/10.1093/toxsci/kft094</u>

10. Dean JL, Zhao QJ, Lambert JC, Hawkins BS, Thomas RS, Wesselkamper SC. Editor's Highlight: Application of gene set enrichment analysis for identification of chemically induced, biologically relevant transcriptomic networks and potential utilization in human health risk Assessment. Toxicol Sci. 2017; 157(1):85-99. <u>http://dx.doi.org/10.1093/toxsci/kfx021</u>

11. Wilson CL, Miller CJ. Simpleaffy: a BioConductor package for Affymetrix quality control and data analysis. Bioinformatics (Oxford, England). 2005; 21(18):3683-3685. http://dx.doi.org/10.1093/bioinformatics/bti605

12. Williams DA. A test for differences between treatment means when several dose levels are compared with a zero dose control. Biometrics. 1971; 27(1):103-117. http://dx.doi.org/10.2307/2528930

13. Williams DA. The comparison of several dose levels with a zero dose control. Biometrics. 1972; 28(2):519-531. <u>http://dx.doi.org/10.2307/2556164</u>

14. Dunnett CW. A multiple comparison procedure for comparing several treatments with a control. Journal of the American Statistical Association. 1955; 50(272):1096-1121. http://dx.doi.org/10.1080/01621459.1955.10501294

15. Shirley E. A non-parametric equivalent of Williams' test for contrasting increasing dose levels of a treatment. Biometrics. 1977; 33(2):386-389. <u>http://dx.doi.org/10.2307/2529789</u>

16. Dunn OJ. Multiple comparisons using rank sums. Technometrics. 1964; 6(3):241-252. http://dx.doi.org/10.1080/00401706.1964.10490181

17. Jonckheere AR. A distribution-free k-sample test against ordered alternatives. Biometrika. 1954; 41(1/2):133-145. <u>http://dx.doi.org/10.2307/2333011</u>

18. Dixon WJ, Massey FJ. Introduction to statistical analysis. New York,: McGraw-Hill; 1951.

19. US EPA. 2012. Benchmark dose technical guidance. Washington, DC: U.S. EPA.

20. Wignall JA, Shapiro AJ, Wright FA, Woodruff TJ, Chiu WA, Guyton KZ, Rusyn I. Standardizing benchmark dose calculations to improve science-based decisions in human health assessments. Environ Health Perspect. 2014; 122(5):499-505. http://dx.doi.org/10.1289/ehp.1307539

21. NTP 2018. NTP research report on National Toxicology Program approach to genomic doseresponse modeling. Research Triangle Park, NC: National Toxicology Program. No. 5. <u>https://doi.org/10.22427/NTP-RR-5</u>

22. Phillips JR, Svoboda DL, Tandon A, Patel S, Sedykh A, Mav D, Kuo B, Yauk CL, Yang L, Thomas RS et al. BMDExpress 2: Enhanced transcriptomic dose-response analysis workflow. Bioinformatics (Oxford, England). 2018. <u>http://dx.doi.org/10.1093/bioinformatics/bty878</u>

23. ATSDR. 2012. Toxicological profile for phosphate ester flame retardants. Atlanta, GA: US Department of Health and Human Services, Public Health Service.

24. Morris PJ, Medina-Cleghorn D, Heslin A, King SM, Orr J, Mulvihill MM, Krauss RM, Nomura DK. Organophosphorus flame retardants inhibit specific liver carboxylesterases and cause serum hypertriglyceridemia. ACS Chem Biol. 2014; 9(5):1097-1103. http://dx.doi.org/10.1021/cb500014

25. Thomas RS, Allen BC, Nong A, Yang L, Bermudez E, Clewell HJ, III, Andersen ME. A method to integrate benchmark dose estimates with genomic data to assess the functional effects of chemical exposure. Toxicological Sciences. 2007; 98(1):240-248. http://dx.doi.org/10.1093/toxsci/kfm092

Appendix A Animal Identifiers and Endpoint Analyses

Animal Number	Group	Dose (mmol/ kg/day)	Dose (mg/kg/ day)	Found Dead	Gavage Accident	Mori- bund	Body Weight	Clinical Obser- vations	Organ Weight	Clinical Chemistry	Hema- tology	Cholin- esterase	T4	Array ID
173	Corn Oil	0	0	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	041-052014-MW_(Rat230_2).CEL
174	Corn Oil	0	0	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	001-051914-MW_(Rat230_2).CEL
184	Corn Oil	0	0	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	073-052714-JAP_(Rat230_2).CEL
210	Corn Oil	0	0	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	009-051914-MW_(Rat230_2).CEL
215	Corn Oil	0	0	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	037-052014-MW_(Rat230_2).CEL
159	TPHP	0.169	55	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	NA
164	TPHP	0.169	55	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	002-051914-MW_(Rat230_2).CEL
168	TPHP	0.169	55	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	NA
171	TPHP	0.169	55	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	038-052014-MW_(Rat230_2).CEL
223	TPHP	0.169	55	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	074-052714-JAP_(Rat230_2).CEL
148	TPHP	0.338	110	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	003-051914-MW_(Rat230_2).CEL
162	TPHP	0.338	110	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	NA
176	TPHP	0.338	110	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	039-052014-MW_(Rat230_2).CEL
179	TPHP	0.338	110	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	075-052714-JAP_(Rat230_2).CEL
194	TPHP	0.338	110	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	NA
160	TPHP	0.675	220	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	004-051914-MW_(Rat230_2).CEL
175	TPHP	0.675	220	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	NA
192	TPHP	0.675	220	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	040-052014-MW_(Rat230_2).CEL
193	TPHP	0.675	220	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	NA
206	TPHP	0.675	220	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	076-052714-JAP_(Rat230_2).CEL

Table A-1. Animal Numbers, Completed Endpoint Analyses, and Microarray Data File Names

Animal Number	Group	Dose (mmol/ kg/day)	Dose (mg/kg/ day)	Found Dead	Gavage Accident	Mori- bund	Body Weight	Clinical Obser- vations	Organ Weight	Clinical Chemistry	Hema- tology	Cholin- esterase	T4	Array ID
186	TPHP	1.35	441	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	005-051914-MW_(Rat230_2).CEL
204	TPHP	1.35	441	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	NA
205	TPHP	1.35	441	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	NA
219	TPHP	1.35	441	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	042-052114-MW_(Rat230_2).CEL
230	TPHP	1.35	441	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	077-052714-JAP_(Rat230_2).CEL
161	TPHP	2.7	881	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	006-051914-MW_(Rat230_2).CEL
191	TPHP	2.7	881	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	NA
200	TPHP	2.7	881	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	043-052114-MW_(Rat230_2).CEL
213	TPHP	2.7	881	No	No	No	Yes	Yes	Yes	Yes	Yes	Yes	Yes	078-052714-JAP_(Rat230_2).CEL

NA = No liver transcriptomics data collected for selected animal.

Appendix B Analytical Chemistry

Analysis	Results					
FTIR and NMR spectroscopy	Consistent with proposed structure and reference spectra					
Ultraviolet/visible spectroscopy	Maxima at λ 267.5, 261.0, and 255.5 nm, consistent with reference spectrum					
Direct infusion mass spectrometry	Confirmation of monoisotopic mass of 327.1 Da representing $[M+H]+$ and consistent with the test article molecular weight					
Elemental analysis	Average % Determined/Theoretical					
Carbon	66.43% 100.26%					
Hydrogen	4.42% 95.46%					
Phosphorus	9.84% 103.69%					
Karl Fischer titration	<0.01% water					
Melting point	48.7° to 49.8 (s)°C					
Differential scanning calorimetry	Purity of $99.72 \pm 0.05(d)\%$					
log P	4.66					
GC purity profile with DB-5 column and flame ionization detection	Purity of 99.35% with four (4) impurities greater than or equal to 0.05%, totaling 0.66%					
Residual solvent content using GC/headspace analysis	No residual solvents present at levels greater than the Class 1 and Class 2 standard mixtures					

Table B-1. Summary of Analytical Results for Triphenyl Phosphate (CAS No. 115-86-6)

GC = gas chromatography; FTIR = Fourier transform infrared; NMR = nuclear magnetic resonance.

Appendix C Toxicology Data Tables

Tables

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Table C-1. I04: Mean Body Weight Summary

Phase	0 mg/kg N = 5	55 mg/kg N = 5	110 mg/kg N = 5	220 mg/kg N = 5	441 mg/kg N = 5	881 mg/kg N = 5
SD0 (g)	258.9 ± 3.6	256.6 ± 5.4	253.5 ± 3.8	261.5 ± 4.2	254.0 ± 6.7	247.8 ± 4.5
SD4 (g)	$276.3 \pm 4.6^{**}$	271.3 ± 5.1	271.9 ± 3.7	278.8 ± 4.7	262.4 ± 8.1	$232.5 \pm 9.4 **$

Data are displayed as mean \pm standard error of the mean.

Statistical analysis performed by Jonckheere's (trend) and Williams' or Dunnett's (pairwise) tests.

Statistical significance for the control group indicates a significant trend test.

**Statistically significant at $p \le 0.01$.

SD = Study Day.

Table C-2. PA06: Organ Weights Summary

Endpoint	0 mg/kg N = 5	55 mg/kg N = 5	110 mg/kg N = 5	220 mg/kg N = 5	441 mg/kg N = 4–5	881 mg/kg N = 5
Terminal Body Weight (g)	$276.3 \pm 4.6^{**}$	271.3 ± 5.1	271.9 ± 3.7	278.8 ± 4.7	262.4 ± 8.1	232.5 ± 9.4**
Brain Weight Absolute (g)	1.73 ± 0.05	1.74 ± 0.03	1.77 ± 0.04	1.81 ± 0.03	1.81 ± 0.03	1.78 ± 0.01
Brain Weight Relative	6.28 ± 0.23**	6.43 ± 0.06	6.52 ± 0.13	6.48 ± 0.03	$7.03 \pm 0.14*$	$7.73 \pm 0.37 **$
Liver Weight Absolute (g)	$11.46 \pm 0.19 **$	11.83 ± 0.56	11.84 ± 0.29	$12.95 \pm 0.48*$	$12.92 \pm 0.64*$	$13.09 \pm 0.38*$
Liver Weight Relative	$41.50 \pm 0.58 **$	43.52 ± 1.41	43.51 ± 0.63	46.42 ± 1.28	$49.14 \pm 1.00 **$	$56.74 \pm 3.22^{**}$

Data are displayed as mean \pm standard error of the mean.

Relative organ weights (organ weight-to-body weight ratios) are given as mg organ weight/g body weight.

Statistical analysis performed by Jonckheere's (trend) and William's or Dunnett's (pairwise) tests.

Statistical significance for the control group indicates a significant trend test.

*Statistically significant at $p \le 0.05$.

Statistical significance for a treatment group indicates a significant pairwise test compared to the vehicle control group.

**Statistically significant at $p \le 0.01$.

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Endpoint Name	0 mg/kg N = 5	55 mg/kg N = 5	110 mg/kg N = 5	220 mg/kg N = 5	441 mg/kg N = 5	881 mg/kg N = 5
Urea Nitrogen (mg/dL)	12.0 ± 0.8	11.6 ± 0.7	12.8 ± 1.3	13.8 ± 0.7	13.2 ± 1.2	30.6 ± 17.2
Creatinine (mg/dL)	0.29 ± 0.01	0.28 ± 0.01	0.29 ± 0.02	0.26 ± 0.01	0.29 ± 0.01	0.4 ± 0.11
Total Protein (g/dL)	5.94 ± 0.13	5.92 ± 0.14	5.9 ± 0.07	6.06 ± 0.08	6.12 ± 0.1	5.84 ± 0.15
Globulin (g/dL)	$2.5\pm0.07^{**}$	2.56 ± 0.07	2.52 ± 0.05	2.66 ± 0.05	2.76 ± 0.06	2.72 ± 0.15
A/G Ratio	$1.38 \pm 0.03 **$	1.31 ± 0.01	1.34 ± 0.02	$1.28\pm0.02*$	$1.22 \pm 0.03^{**}$	$1.16 \pm 0.07 **$
Albumin (g/dL)	$3.44\pm0.07*$	3.36 ± 0.07	3.38 ± 0.04	3.4 ± 0.04	3.36 ± 0.06	$3.12\pm0.06*$
Cholesterol (mg/dL)	$101.0 \pm 5.8^{**}$	115.0 ± 6.5	122.4 ± 6.2	$127.8\pm6.3^*$	$145.4 \pm 7.3^{**}$	$170.0 \pm 19.4 ^{**}$
Triglyceride (mg/dL)	$54.0\pm6.0*$	53.8 ± 1.4	61.6 ± 6.8	74.0 ± 6.1	53.0 ± 5.3	140.6 ± 59.1
LDL Cholesterol (mg/dL)	$21.8 \pm 0.7 **$	21.8 ± 1.1	23.2 ± 0.9	23.8 ± 1.5	23.4 ± 1.1	$32.8 \pm 6.4*$
HDL Cholesterol (mg/dL)	46.0 ± 2.9**	54.6 ± 2.7	$56.8\pm2.8*$	60.6 ± 2.9**	70.6 ± 3.2**	78.2 ± 5.1 **
Alanine Aminotransferase (IU/L)	61.4 ± 4.46	77.4 ± 6.55	63.8 ± 5.54	65.0 ± 5.37	65.6 ± 6.25	116.6 ± 32.1
Aspartate Aminotransferase (U/L)	107.4 ± 10.21	94.6 ± 10.38	97.8 ± 9.82	73.0 ± 2.17*	80.8 ± 9.89	150.8 ± 46.99
Sorbitol Dehydrogenase (IU/L)	10.4 ± 0.7	11.8 ± 1.1	12.7 ± 1.6	10.4 ± 0.7	10.8 ± 0.7	11.9 ± 1.0
Bile Salt/Acids (µmol/L)	57.4 ± 8.8	39.5 ± 5.5	26.6 ± 3.2*	37.2 ± 6.3	36.2 ± 10.2	38.9 ± 5.5

Table C-3. PA41: Clinical Chemistry Summary

Statistical analysis performed by Jonckheere's (trend) and Shirley's or Dunn's (pairwise) tests.

Statistical significance for the control group indicates a significant trend test.

*Statistically significant at $p \le 0.05$.

Statistical significance for a treatment group indicates a significant pairwise test compared to the vehicle control group. **Statistically significant at $p \le 0.01$.

Table C-4. R07: Hormone Summary

Endpoint Name	0 mg/kg N = 5	55 mg/kg N = 5	110 mg/kg N = 5	220 mg/kg N = 5	441 mg/kg N = 5	881 mg/kg N = 5
Free Thyroxine (µg/dL)	5.122 ± 0.227**	5.114 ± 0.196	4.678 ± 0.304	$4.274 \pm 0.263*$	$4.066 \pm 0.331*$	1.870 ± 0.429**
Cholinesterase (IU/L)	$277.4 \pm 16.6^{**}$	$178.4 \pm 8.0 **$	$162.6 \pm 7.6^{**}$	$116.6 \pm 5.8 **$	$108.8 \pm 8.3 **$	$82.8 \pm 2.1 **$

Data are displayed as mean \pm standard error of the mean.

Statistical analysis performed by Jonckheere's (trend) and Shirley's or Dunn's (pairwise) tests.

Statistical significance for a treatment group indicates a significant pairwise test compared to the vehicle control group.

Statistical significance for the control group indicates a significant trend test.

*Statistically significant at $p \le 0.05$.

**Statistically significant at $p \le 0.01$.

Appendix D Transcriptomic Quality Control and Additional Data Analysis



D.1 Gene Expression Quality Control

Figure D-1. A Principal Component (PCA) of the RMA-Normalized Data

The PCA plot enables three-dimensional visualization of global transcriptional changes and the divergence of transcript expression from individual animals (dots) within each dose group (designated by color). Dots that are spatially closer to each other indicate more similarity in global expression profiles; dots that are farther apart indicate dissimilarity in global expression profiles for those animals. Lighter color (fogging) indicates that a data point is farther back on the z-plane [principal component (PC) #1]. [Note: 440 mg/kg dose is equivalent to 441 mg/kg.]



D.2 Additional Data Analysis

Figure D-2. An Alternative View of the Principal Component (PCA) of the RMA-Normalized Data

This alternative view of the PCA plot enables visualization of global transcriptional changes in two dimensions, with each plot showing a different angle, based on the principle components plotted. Global transcript data are shown for individual animals (dots) within each dose group (designated by color). Dots that are spatially closer to each other indicate more similarity in global expression profiles; dots that are farther apart indicate dissimilarity in global expression profiles for those animals. [Note: 440 mg/kg dose is equivalent to 441 mg/kg.]

Appendix E Apical Endpoint Quality Control and Additional Data Analysis

Rule	Criteria for "Viable"	Numerical Threshold (N)	Bin Placement for Rule Failure
BMD existence	A BMD exists.	N/A	Failure
BMD_L existence	A BMD _L exists.	N/A	Failure
AIC existence	An AIC exists.	N/A	Failure
Residual of interest existence	The residual at the dose group closest to the BMD (i.e., the residual of interest) exists.	N/A	Failure
Constant variance model selection	The constant variance model is appropriate (BMDS Test 2 p-value $>$ N).	0.1	Nonviable
Variable variance model selection	The variable variance model is appropriate (BMDS Test 2 failed and BMDS Test 3 p-value > N).	0.1	Nonviable
Global goodness of fit	The mean model fits the data means sufficiently well (BMDS 2.7 Test 4 p-value $>$ N).	0.1	Nonviable
BMD-to-BMD _L ratio	The ratio of BMD to BMD_L is not large $(BMD/BMDL < N)$.	20	Nonviable
$High BMD_L$	The BMD_L is $ times higher than the maximum dose.$	1	Nonviable
Low BMD	The BMD is <n dose.<="" lower="" minimum="" nonzero="" td="" than="" the="" times=""><td>10</td><td>Nonviable</td></n>	10	Nonviable
Low BMD _L	The BMD _L is <n dose.<="" lower="" minimum="" nonzero="" td="" than="" the="" times=""><td>10</td><td>Nonviable</td></n>	10	Nonviable
Control residual	The residual at control is small (residual < N).	2	Nonviable
Control standard deviation	The modeled standard deviation is similar to the actual (<n different).<="" td="" times=""><td>1.5</td><td>Nonviable</td></n>	1.5	Nonviable
Residual of interest	The residual at the dose group closest to the BMD (i.e., the residual of interest) is small (residual < N).	2	Nonviable
No warnings reported	No warnings in the BMD model system were reported.	N/A	Viable

Table E-1. Benchmark Dose Model Recommendation/Selection Rules for Apical Endpoints

 $\overline{AIC} = Akaike information criterion; BMD = benchmark dose; BMD_L = benchmark dose lower confidence limit; N/A = not applicable.$



Figure E-1. Benchmark Dose Model Recommendation/Selection Methodology for Automated Benchmark Dose Execution of Apical Endpoints

Source: Figure from Wignall et al. 2014²⁰.



Figure E-2. Benchmark Dose Model Recommendation/Selection Methodology for Benchmark Dose Execution of Gene Sets with Expression Changes Enacted by Chemical Exposure

Adapted from Thomas et al. 25.

Abbreviations: AIC = Akaike information criterion; BMD = benchmark dose; $BMD_L =$ benchmark dose lower confidence limit; $BMD_U = BMD$ upper confidence limit; GGOF = global goodness of fit; GO = Gene Ontogeny.



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